



# UPDATE - T&E's analysis of electric car lifecycle CO<sub>2</sub> emissions

June 2022

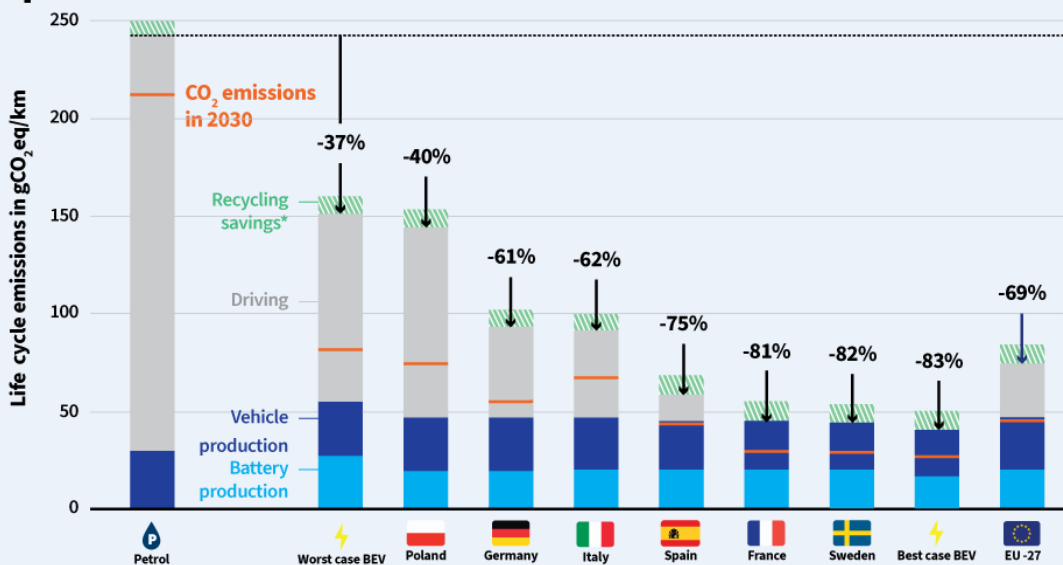
## Summary

Following the entry into force of the 2020/21 EU car CO<sub>2</sub> standards, battery electric cars (BEVs) have stepped into the mass market much faster than previously expected with one out of ten cars sold in the European area being battery electric in 2021. In parallel, the sales of plug-in hybrid vehicles (PHEVs) and full hybrids (HEVs) also grew in recent years and there are still some misconceptions about how clean these powertrains are compared to BEVs.

To bring clarity to the subject, T&E has updated its [online tool](#) comparing the lifecycle CO<sub>2</sub> emissions of different powertrains of vehicles sold in 2022 and 2030 and has added PHEVs and hybrids. This lifecycle assessment (LCA) is based on the latest available real world direct emissions data, the real world share of distance driven electrically by PHEVs, as well as the latest data and forecasts of the carbon intensity of the European electricity grid. Technologies such as solid-state batteries and synthetic fuels (“e-fuels”) which may be available in limited quantities from 2030 have also been added to the tool.



## Average EU electric cars are more than 3 times cleaner than petrol cars

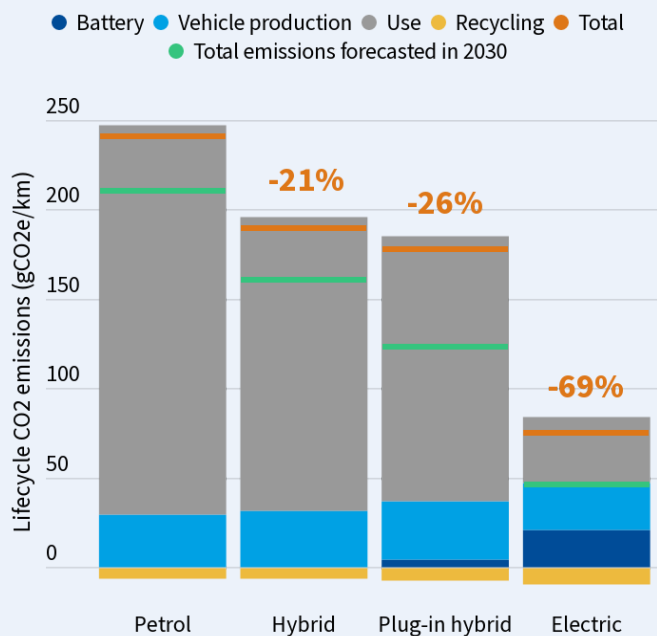


\*Recycling savings apply to the production phase

Note: T&E LCA analysis of a medium-sized car, battery assumed to be produced with the EU27 average grid. Worst case: battery produced in China, charging in Poland. Best case: battery produced in Sweden, charging in Sweden.

This new analysis reveals that the average European BEV is more than three times cleaner than equivalent petrol cars bought in 2022 (-69% CO<sub>2</sub> emissions). In the best case, where the battery production and the charging use the cleanest electricity grid, a medium-sized BEV is nearly six times cleaner (-83%). On the other hand, in the worst case, where the battery is produced in China and the car is charged in Poland, a BEVs is still 37% cleaner than petrol. Based on an average battery, an electric car driven in Germany or Italy is slightly more than 2.5 times cleaner than a conventional car (-61% and -62%), while BEVs driven in France and Sweden have an impact more than five times lower (-81% and -82%). BEVs bought in 2030 will fully benefit from the cleaner electricity grid (renewable electricity is expected to make up 62% of the EU electricity in 2030) and are expected to be 4.6 times cleaner than petrol.

## Lifecycle emissions per powertrain



**Source:** T&E LCA analysis of a medium-sized car, battery assumed to be produced with the EU27 average grid, BEV/PHEV charging with the EU27 average grid

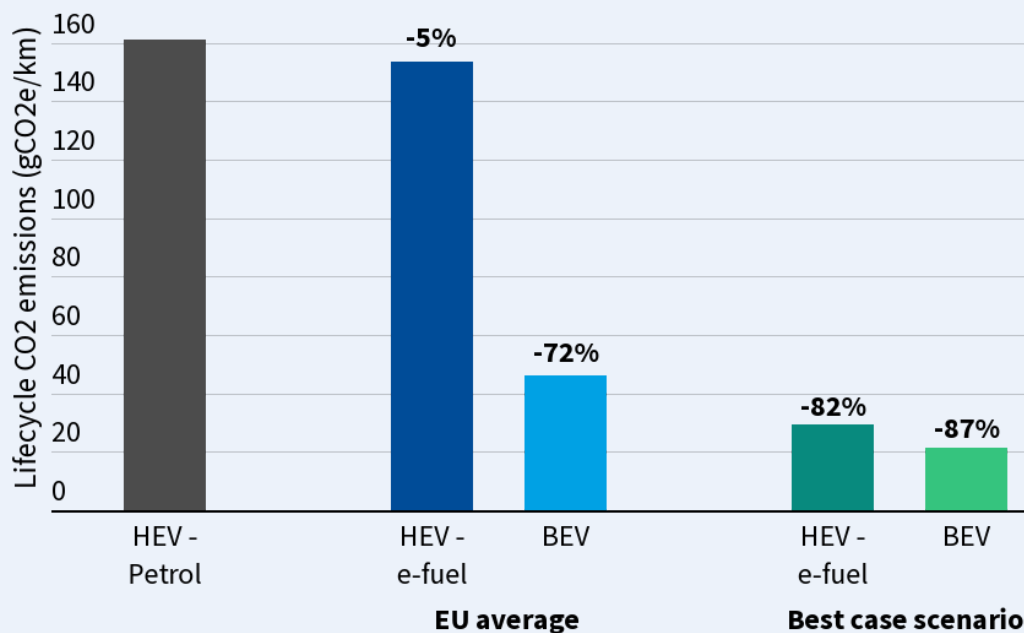
TRANSPORT & ENVIRONMENT [transportenvironment.org](https://transportenvironment.org)

Despite their green credentials, HEVs' and PHEVs' lifecycle emissions are much closer to polluting conventional petrol cars than to BEVs. The results show that HEVs only achieve a 21% reduction in LCA emissions compared to an equivalent petrol car while PHEV improvements are limited to 26%. As a result PHEVs are more than two times more polluting than BEVs. This poor performance of PHEVs is caused by the fact that in real world operations, only about 38% of the kilometres are driven in electric mode (significantly less than is assumed by the official EU emission test).

PHEVs bought in 2030 are expected to be driven more on the electric powertrain, but thanks to the rapidly decarbonising grid, BEVs would be 2.7 times cleaner than PHEVs by this point. Policy makers must not make room in EU climate policy for this sub-optimal technology that offers such limited CO<sub>2</sub> benefits compared to conventional cars.

The results also demonstrate the limited CO<sub>2</sub> reduction potential of synthetic fuels. For example, running a petrol hybrid car on a blend of synthetic and conventional petrol would only reduce emissions by 5%-7% for cars bought in 2030. The blend modelled by T&E is based on forecasts done by the fuel industry where 0.4% of the fuel available at the pump would be synthetic fuel in 2030 (3% in 2035, 16% in 2040). Looking at a more theoretical scenario where a car is powered entirely by e-fuels, the emissions would be reduced more significantly: -82% if the e-fuel is made from 100% renewables and -53% if it is made using the RED II fuel sustainability criteria (where 15%-30% of the energy required for the e-fuel production could still come from non-renewable sources). Nonetheless, in the best case scenario, a BEV would still be 27% cleaner than the best case for combustion engines (HEV running on 100% e-fuel made from 100% renewable energy) mainly due to the low efficiency of the e-fuel production process. Nevertheless, results with 100% e-fuel could only be applicable to a marginal part of the car fleet as only about 3% of internal combustion engine vehicles (ICEs) could run on 100% e-fuel in 2035 (according to the optimistic forecast from the fuels industry). With such strong supply constraints, e-fuels will be better used in hard to decarbonise sectors such as aviation and shipping, where no better alternatives exist at the moment.

## Electric vehicles are cleaner than cars running on e-fuel



**Average use:** e-fuel produced according to the regulation criteria (REDII) and blended with petrol based on industry fuel production forecast. Average EU grid used for battery production and charging, and average supply-chain impacts.

**Best case scenario:** pure e-fuel produced from 100% renewable. Swedish electricity grid used for battery production and charging, and low supply-chain impacts.

**BEVs are cleaner:** even under the best case scenario, the production of e-fuel has low energy efficiency and requires a greater amount of energy, meaning it has higher lifecycle emissions due to indirect emissions from the electricity infrastructure.

**Source:** T&E LCA analysis of a medium-sized car bought in 2030

In 2030, the LCA emissions of BEVs can be even further reduced with different levers. First, the use of cleaner electricity for battery production would enable a 5% savings on the BEV lifecycle emissions. Second, as battery supply chains are also getting cleaner, best-in-class supply chains used for the sourcing of lithium, nickel or graphite can further decrease BEV lifecycle emissions by a further 13%. Therefore, BEVs bought in 2030 have the highest potential for CO<sub>2</sub> emissions reductions over their lifecycle.

This latest evidence clearly highlights that battery electric cars remain the most promising and mature technological solution to accelerate the transition to carbon neutrality. Given the urgency to decarbonise all sectors of our economy, the EU should not delay this transition with sub-optimal technology such as hybrids nor with extremely limited, expensive and inefficient synthetic fuels. The phase out of all internal combustion engines by 2035 without the option for carmakers to purchase credits for e-fuels, is the only way to stay in line with the EU's Green Deal ambitions.

# 1. Methodology

The methodology for T&E’s electric vehicle (EV) Life Cycle Assessment (LCA) of CO<sub>2</sub> emissions was described in April 2020 in the report entitled [‘How clean are electric cars?’](#). The following sections explain methodological modifications that were included in April 2022. (For any data or method not mentioned in this briefing, please refer to the April 2020 report.)

## 1.1 Electricity sources emissions factors and electricity mixes

In the previous version of this LCA analysis, T&E’s calculation of the carbon intensity of a country’s electricity grid was made using of a bottom-up calculation based on the electricity generation mix evolution forecasted in 2020 and the Intergovernmental Panel on Climate Change (IPCC) global emissions factors for different energy sources. In this latest version, new EU-specific emissions factors derived from the 2021 UNECE report<sup>1</sup> have been included. The electricity generation mix is now based on ENTSO-E’s TYNDP 2022 Draft Scenario Report<sup>2</sup>.

### Life-cycle carbon intensity of electricity sources

To replace IPCC emissions factors that were defined in the Fifth Assessment Report (2014), T&E derived new emissions factors from a recent LCA analysis published by UNECE in 2021. This study (based on more recent data sources) provides results customised to the European region instead of global averages. The following emissions factors were included in T&E analysis<sup>3</sup> based on hypotheses described in Annex 1. Compared to the previously used 2014 IPCC data, results derived from UNECE lead to a lower carbon intensity for hydroelectric, solar and nuclear power.

**Table 1 – Life-cycle carbon intensity of electricity sources**

LCA CO <sub>2</sub> eq. emissions <sup>4</sup>	Coal	Gas	Solar PV	Offshore Wind	Hydro	Onshore Wind	Nuclear
(gCO <sub>2</sub> e/kWh)	997	434	34	14	11	12	5

### EU countries electricity grid

For each country, the lifecycle carbon intensity of electricity was assumed to evolve linearly between 2020 and 2025, between 2025 and 2030 and between 2030 and 2040. Data from 2020 are historic data from Ember<sup>5</sup> while forecasted data are from ENTSO-E’s TYNDP 2022 Draft Scenario Report. ENTSO-E’s scenarios depict European energy futures up to 2050. The scenarios used are the ‘National Trends’ for

<sup>1</sup> <https://unece.org/sites/default/files/2021-10/LCA-2.pdf>

<sup>2</sup> <https://2022.entsos-tyndp-scenarios.eu/>

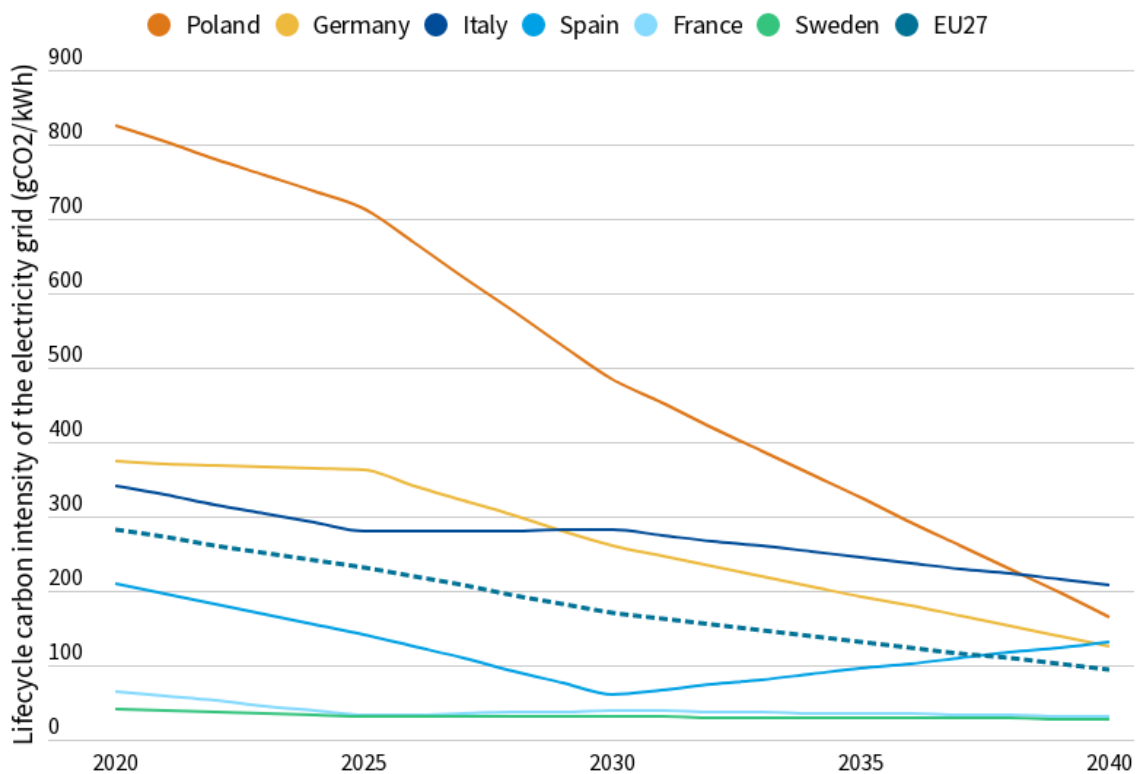
<sup>3</sup> Other electricity sources emissions are unchanged compared to T&E 2020 report

<sup>4</sup> Radiative forcing as global warming potential, integrated over 100 years (GWP100)

<sup>5</sup> <https://ember-climate.orgCO2e/insights/research/eu-power-sector-2020/>

2025 and the ‘Global Ambition’ for 2030 and 2040, taken from the official European network of Transmission System Operators (TSOs), ENTSO-E’s scenarios, and particularly the ‘Global Ambition’ scenarios. We consider these forecasts to be the most credible pathways, best reflecting the European net-zero carbon ambition for 2050.

Taking into account the above update of emissions factors and country generation mixes, the calculated EU27 average life-cycle carbon intensity of the electricity grid is 261 gCO<sub>2</sub>e/kWh<sup>6</sup> in 2022, 171 gCO<sub>2</sub>e/kWh in 2030 and 93 gCO<sub>2</sub>e/kWh in 2040. Vehicles sold in 2022 are used as the default option of the LCA model whereas the previous version of the model used vehicles sold in 2020 with 319 gCO<sub>2</sub>e/kWh taken as the EU27 average carbon intensity of the electricity grid (based on IPCC emissions factors and a previous estimate of the EU grid mix). The methodology accounts for yearly change in carbon intensity over the lifetime of the vehicles. This update leads to a 18% improvement of the electricity carbon intensity in the default option, which explains the improvement of BEVs lifecycle emissions referred to in section 2.1. Results for key EU countries are shown in the figure below:



**Source:** T&E modelling based on future electricity generation from ENTSOE TYNDP 2022 Draft Scenario Report and lifecycle emissions factors derived from UNECE LCA report (2021)

**Figure 1 – BEV lifecycle emissions savings in T&E updated model**

<sup>6</sup> Lifecycle carbon intensity including transport and distribution losses

## 1.2 Fuel consumption and vehicle data

### Fuel consumption used for the 2022 baseline option

The latest real-world fuel consumption data was extracted from the spritmonitor.de database, a German website that collects information about the fuel consumption of vehicles under real-life conditions. In addition to petrol and diesel cars, full hybrid vehicles (HEVs<sup>7</sup>) and plug-in hybrids have been added to the analysis. In order to gather representative averages for each vehicle segment, a sufficient number of models were included to cover at least 70% of sales for each segment<sup>8</sup>, related to each powertrain. Overall, the extracted data amounts to 74% of petrol car sales, 86% of diesel sales and 86% of HEV sales (EU27 sales figures from 2021).

Regarding plug-in hybrid electric vehicles (PHEVs), the real world fuel and electricity consumption depends on the share of distance driven on electricity (utility factor) or the use pattern for the car (private or company car drivers don't have the same utility factor). T&E analyses of PHEV consumption data are detailed in Annex 2. Real-world consumption is estimated from the carmaker's fuel and electricity consumption and the equivalent all-electric range values measured on the Worldwide Harmonised Light Vehicles Test Procedure (WLTP), a type-approval test. Correction factors are used to estimate real-world fuel consumption from WLTP values and realistic utility factors (UF) are included to be representative of the average user of a PHEV. These realistic utility factors were calculated by Fraunhofer ISI<sup>9</sup> based on the average range driving on electricity by real PHEV users.

The fuel and electricity consumption, as well as the battery capacity used in the model are provided in the table below:

**Table 2 – Powertrains parameters in the 2022 model**

Segment	Gasoline (L/100km)	Diesel (L/100km)	HEV (L/100km)	PHEV (L/100km)	PHEV (kWh/100km)	PHEV battery capacity (kWh)	BEV (kWh/100km)	BEV battery capacity (kWh)
Small	6.5	5.4	4.8	NA <sup>10</sup>	NA	NA	16.0	45
Medium	7.5	6.2	5.7	4.5	7.8	12	17.5	60
Large	8.8	6.7	6.6	5.0	8.4	13	19.0	75
Executive	8.5	6.9	7.0	5.2	11.4	15	20.5	90

<sup>7</sup> In this analysis, HEVs refer only to full hybrid vehicles as mild hybrids are grouped with conventional ICE.

<sup>8</sup> Based on T&E analysis of 2021 registration data in EU27 provided by Dataforce

<sup>9</sup> [https://www.isi.fraunhofer.de/content/dam/isi/dokumente/cce/2021/BMU\\_Kurzpapier\\_UF\\_final.pdf](https://www.isi.fraunhofer.de/content/dam/isi/dokumente/cce/2021/BMU_Kurzpapier_UF_final.pdf)

<sup>10</sup> Very few PHEV models are sold in segment B (small) and the methodology defined to assess real world consumption is not applicable to individual models but only to an average model. For that reason, results would not be representative in this segment and thus they have been omitted.

## 2030 fuel consumption forecast

A forecast of emissions by powertrain based on latest industry trends was defined by T&E in the [Electric car boom at risk](#) report. Based on this analysis, the following assumptions are used to calculate expected emission reductions by powertrain:

- Pure ICE fuel consumption by body type, such as hatchbacks and SUVs, is assumed to follow a downward trend of 1.5% each year between 2021 and 2025 and is expected to stagnate as from 2025. Mild hybrid fuel consumption is assumed to follow a similar downward trend of 1.5% each year between 2021 and 2025, and then 1% each year from 2026. The growth of the SUV share is expected to follow the historical trend and grow from 40% in 2020 to 63% in 2030. However, this anticipated SUV growth would slow the emissions reductions. Moreover, the sales of mild hybrids among ICE is estimated to grow from 11% in 2020 to 91% in 2030. Together with improvements in fuel consumption, these changes in the mild hybrid share would lead to an average ICE fuel consumption decrease of 1.1% every year from 2022 to 2030.
- Full hybrid fuel consumption is assumed to decrease by 1.5% each year between 2021 and 2025 and 1% each year from 2026.
- BEV electricity consumption is assumed to decrease by -0.5% each year.
- Regarding PHEVs, fuel consumption is expected to decrease in line with HEVs and the electricity consumption trend should decrease, in line with BEVs. However, the utility factor of PHEVs is likely to increase since their range is planned to rise. Another factor is that the share of private users (who have a higher electric driving share vs. company car users) is also expected to increase, see Annex 2 for all assumptions. The increase of PHEV battery capacity is based on industry trends analysed by T&E in the [Promises, but no plans](#) report.

The forecasted values are provided in the following table:

**Table 3 – Powertrains: parameters in the 2030 model**

Segment	Gasoline (L/100km)	Diesel (L/100km)	HEV (L/100km)	BEV (kWh/100km)	PHEV (L/100km)	PHEV (kWh/100km)	PHEV Battery capacity (kWh)
Small	5.9	4.8	4.3	15.3	NA	NA	NA
Medium	6.8	5.7	5.1	16.7	3.2	10.6	17.1
Large	8.0	6.1	5.9	18.2	3.7	12.3	21.7
Executive	7.7	6.2	6.3	19.6	3.6	16.6	23.0



### **Emissions from production of glider and powertrain**

From the analysis in the 2020 Ricardo LCA report<sup>11</sup>, T&E estimates that the production of HEV glider and powertrain is 5.9% more carbon intensive (battery included<sup>12</sup>) than its petrol equivalent (provided in T&E 2020 LCA report), and that PHEV glider production is 9.6% more carbon intensive (battery excluded<sup>13</sup>).

## **1.3 Carbon footprint of battery production**

T&E's battery carbon footprint model was calibrated based on results from an LCA report<sup>14</sup> that T&E commissioned from Minviro. In this report, Minviro calculated a 77 kgCO<sub>2</sub>e/kWh carbon footprint for a NMC-811 lithium-ion battery (LIB) produced in 2021 with the average EU27 electricity grid. Based on this data, T&E estimated that an NMC-622 battery produced with the EU grid in 2022 would have a 78 kgCO<sub>2</sub>e/kWh carbon footprint. In 2030, we expect the footprint of an NMC-811 battery to decrease to 55 kgCO<sub>2</sub>e/kWh based on the grid carbon intensity forecast described in section 1.1.

Minviro analysed the impact of using best-in-class material sourcing for key materials of an NMC battery, for instance by using lithium from geothermal sources, nickel from a bioleaching process and synthetic graphite produced in Europe. By integrating Minviro results into the model, it was estimated that an NMC-622 battery produced with the 2022 EU27 electricity grid and a low impact supply chain would have a 48 kgCO<sub>2</sub>e/kWh carbon footprint.

Moreover, Minviro estimated the carbon footprint of future battery technologies such as solid state batteries (SSB). The SSB formulation with the lowest carbon footprint would be based on a NMC-811 cathode, lithium metal anode and an oxide-based solid electrolyte. T&E estimated that this formulation would achieve a carbon footprint of 43 kgCO<sub>2</sub>e/kWh with the 2030 electricity EU grid.

Each battery type (LIB or SSB) and sourcing option (standard or low impact supply chain) was modelled based on different production locations detailed in Annex 3, so as to assess the impact of the local electricity grid (2022 or 2030 electricity grid) used for manufacturing processes.

The different battery options are presented in the table below:

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<sup>11</sup> <https://op.europa.eu/en/publication-detail/-/publication/1f494180-bc0e-11ea-811c-01aa75ed71a1>

<sup>12</sup> Emissions from HEV battery production are expected to be relatively small, their impact is not modelled explicitly.

<sup>13</sup> PHEV battery capacity is expected to increase significantly in 2030, therefore PHEV battery carbon footprint is modelled explicitly based on battery capacity (same hypotheses as BEV described in section 1.3).

<sup>14</sup> Report not yet published at the time of writing

**Table 4 – Carbon content of batteries (kgCO<sub>2</sub>e/kWh)**

		Li-ion battery (NMC622 in 2022, NMC811 in 2030)		Solid state battery (NMC 811 - oxide)	
Year	Battery production	Standard	Low supply chain impact	Standard	Low supply chain impact
2022	EU average	78	48		
	Sweden	64	35		
	Germany	85	55		
	China	105	75		
2030	EU average	55	33	43	33
	Sweden	47	25	37	27
	Germany	61	39	47	37
	China	81	59	63	53

According to T&E tracking of battery factory plans in Europe, Germany could produce around 40% of European batteries in 2030. The German grid is expected to be 56% more carbon intensive than the European average in 2030. Therefore, Germany is now included in the tool to study sensitivity for a higher-than-average carbon intensity of the grid.

## 1.4 Greenhouse gas emission reductions from e-fuels

### Production of e-fuels

The lifecycle analysis of ICEs powered by e-fuels<sup>15</sup> differs from conventional fuel as the e-fuel well-to-wheel (WTW) emissions are lower thanks to the use of renewable energy in the fuel production processes. In this study, two e-fuel production alternatives are considered: (1) e-fuel produced under the RED II sustainability criteria (see further details below) as the baseline option: and (2) e-fuel produced using 100% renewable electricity.

In option (1), the Renewable Energy Directive (RED II) sustainability criteria is used to determine how much renewable electricity is necessary to produce the e-fuel. The RED outlines a regulatory framework

<sup>15</sup> Synthetic fuels (or e-fuels) are produced by combining hydrogen and carbon in order to create a hydrocarbon (like petrol or diesel) which can be used to propel a conventional petrol or diesel vehicle. The hydrogen can be produced via electrolysis by splitting water into hydrogen and oxygen molecules while the carbon can be obtained via direct carbon capture. Only synthetic fuels produced from electricity are considered in this paper. Biofuels are not included in this analysis.

to ensure the sustainability of so-called renewable fuels of non-biological origin (RFNBOs) by requiring at least 70% greenhouse gas<sup>16</sup> savings compared to their fossil fuel equivalent (using a WTW accounting). In effect, this implies that a high share of renewable electricity will be needed to meet this threshold. Around 86% of the electricity will have to come from renewables if a combination of renewable electricity and gas is used.

In the 100% renewable electricity scenario, T&E assumes that the e-fuel is produced with a mix of different renewable sources which is derived from the BloombergNEF<sup>17</sup> electrolyser forecast: 34% offshore wind, 31% onshore wind, 28% solar and 7% hydro. The lifecycle CO<sub>2</sub> emissions from fuel production are calculated based on this electricity carbon intensity and an overall 55% e-fuel production efficiency<sup>18</sup>. Based on the expected electrolyser capacity in Europe in 2030 and their electricity source, T&E estimated that the carbon intensity of the electricity used for the e-fuel production would reach 19 gCO<sub>2</sub>e/kWh (assuming that the electricity source used for electrolysis is representative for the electricity used for the e-fuel production). This lifecycle emissions analysis thus provides a complete picture, including indirect upstream emissions (e.g. fuel production and refining) as well as indirect emissions from the necessary infrastructure (e.g. emissions for the production of renewables infrastructure such as solar panels and wind turbines). Since these upstream sources are not zero emission even under the 100% renewable scenario e-fuel CO<sub>2</sub> intensity is 19 gCO<sub>2</sub>e/kWh.

### Blending of e-fuel

E-fuels are drop-in fuels, which means they can substitute petrol and diesel in conventional cars and can be used under different fuel blends ratios, according to industry with a blend ratio anywhere from 0-100%<sup>19</sup>. T&E defined two e-fuel blending scenarios for petrol cars:

- Industry e-fuel blending scenario: This is the baseline scenario, based on the assumption that the available e-fuel is sold blended with petrol. Under this scenario it can be expected that an increased quantity of e-fuels would be gradually added to conventional fuels as the e-fuel industry ramped-up its production over time. This blending scenario is based on the Concawe's (division of the European Petroleum Refiners Association) 'Alternative 1.5°C' scenario<sup>20</sup> where e-fuels would be used in all transport modes including road transport. Concawe forecasts both the overall liquid fuel demand and the installation of new units for e-fuel production. From this, they allocated part of the e-fuel production between transport modes to achieve an accelerated substitution of fossil-derived fuels. Concawe assumed that 1 Mtoe of e-fuel would be attributed to

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<sup>16</sup> In this analysis T&E assumes that the 70% GHG reduction criteria for e-fuels is calculated based on WTW emissions from energy sources (i.e. not lifecycle emissions, which means that renewables are counted as zero - no infrastructure related emissions). In this methodology, direct air capture of CO<sub>2</sub> is assumed and the different point source possible is not considered (only the energy used to perform carbon capture and utilisation is accounted for).

<sup>17</sup> BloombergNEF tracks the announced electrolyser nameplate capacity and the electricity source type of these projects.

<sup>18</sup>

[https://www.transportenvironment.org/sites/te/files/publications/2020\\_12\\_Briefing\\_feasibility\\_study\\_renewables\\_decarbonisation.pdf](https://www.transportenvironment.org/sites/te/files/publications/2020_12_Briefing_feasibility_study_renewables_decarbonisation.pdf)

<sup>19</sup> <https://www.efuel-alliance.eu/efuels/costs-outlook>

<sup>20</sup> Scenario 1 from [https://www.concawe.eu/wp-content/uploads/Rpt\\_21-7.pdf](https://www.concawe.eu/wp-content/uploads/Rpt_21-7.pdf)

road transport in 2030, and that this amount would increase to 6 Mtoe in 2035, 21 Mtoe in 2040 and 46 Mtoe in 2050. T&E assumed that the average blending of e-fuels would be aligned with the e-fuel share of all liquid fuel demand in road transport from the Concawe study. This would lead to an average blending of 0.4% e-fuels in 2030, then successively 3% in 2035, 16% in 2040 and 50% in 2050. Over the lifetime of a car bought in 2030, which stays on the road for 15 years, the average lifetime e-fuel blend would thus be limited to around 10%.

- 100% e-fuel: The scenario assumes a direct use of the e-fuel in a car (100% e-fuel). This scenario is highly theoretical, since the e-fuel supply dedicated to cars is expected to be very limited in the future and to be blended with conventional petrol (see baseline scenario). The e-fuel supply allocated to road transport is expected to cover a minor part of the demand. Without blending, only about 3% of the non-electric car fleet in 2035 would be able to run on 100% e-fuel.

Both scenarios assume that part of EU e-fuel production -and the green hydrogen used in its production- are allocated to road transport. Consequently, this would divert production and supply of synthetic fuels or green hydrogen from other sectors such as aviation, shipping or heavy industry where they are most needed as no better decarbonisation solutions exist for these sectors. Both scenarios are highly inefficient and the 100% e-fuel scenario is given for indicative purposes but it cannot be scaled to cover a significant part of the car fleet.

## 1.5 Estimate of recycling credits

While recycling processes consume energy and initially increase emissions, they ultimately reduce the need for new raw materials to be extracted from the ground for future vehicles. Results from Ricardo's<sup>21</sup> LCA (commissioned by the EU Commission) shows that recycling enables a reduction of the production carbon footprint of a car of about 19% in 2020 and 22% in 2030. Recycling credits were added in the update of T&E's methodology following the same methodology and findings for recycling credits. These credits are negative GHG emissions given at the end of life to account for the benefits accrued from avoiding raw material extraction. CO<sub>2</sub> credits are given based on the difference between the recycled content of materials used for production and the total recycling rate at the end of the vehicle's life. This conservative method implies that credits are only given when the recycling rate significantly exceeds the content of secondary material used in input. Based on the Ricardo study's results, T&E applied a simplified methodology to account for recycling credits for all powertrains which are estimated to be about 19% of the production carbon footprint in 2022, and 22% in 2030. This estimate, utilising a relative value, has enabled us to scale the Ricardo results and therefore to have a consistent order of magnitude applicable to T&E's assumptions and methodology for the production carbon footprint.

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<sup>21</sup> Ricardo's recycling methodology is based on the Product Environmental Footprint 'Circular Footprint Formula' methodology for vehicle recycling and GREET data for battery recycling. <https://op.europa.eu/en/publication-detail/-/publication/1f494180-bc0e-11ea-811c-01aa75ed71a1>

## 2. Results

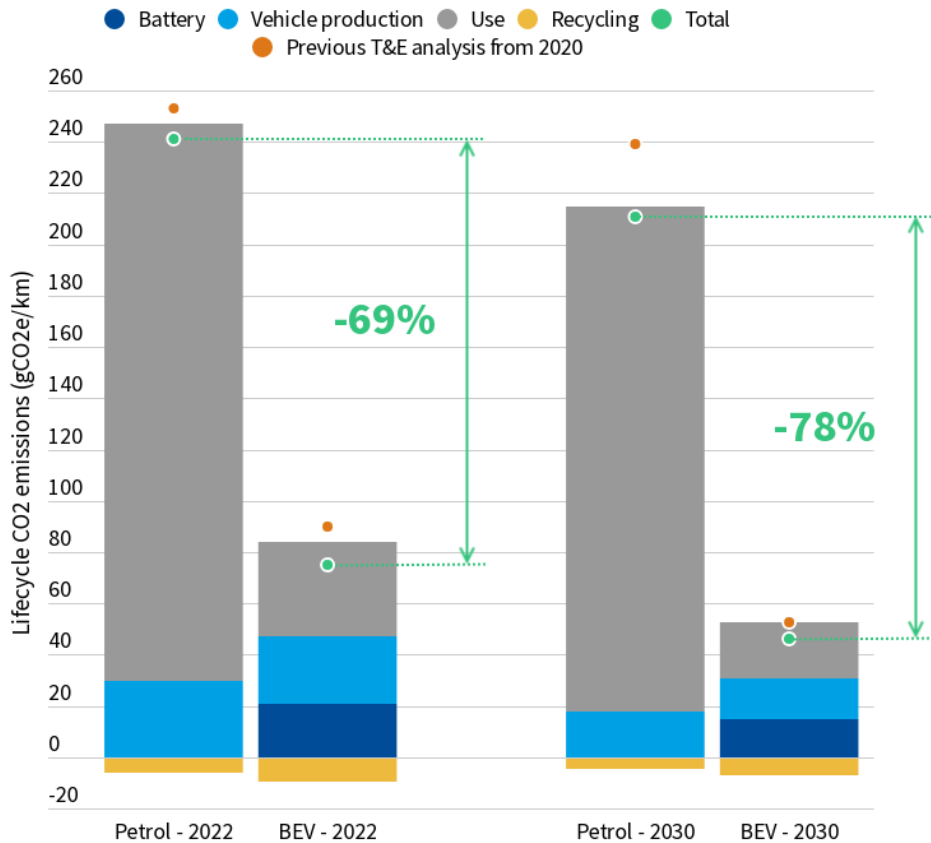
### 2.1 Comparison with T&E's EV LCA from 2020

For cars sold in 2022, a medium-sized EU-average electric car emits 75 gCO<sub>2</sub>e/km over its lifetime, while the lifetime emissions of a petrol car emits 241 gCO<sub>2</sub>e/km, including all upstream emissions and end-of-life credits. This shows that BEVs emit about 3.2 times less than petrol cars. These results are slightly better than what was previously shown by T&E in 2020 (BEV were found to emit 90 gCO<sub>2</sub>e/km over the lifetime, or 2.8 less than petrol). The changes in the results are primarily explained by improvements in the European electricity grid since 2020, the update of the electricity sources emission factors and the addition of recycling credits in the model. For large and executive categories, average EU electric cars are between 3.4 and 3.6 times better than the petrol equivalent.

When the battery is produced with the cleanest electricity grid, the impact of electric cars decreases to 71 gCO<sub>2</sub>e/km or 3.4 times less than petrol. In the best scenario, if the electric car runs mostly on renewable electricity (e.g. Sweden hydro power) and the battery was produced with the cleanest electricity grid, then the GHG impact decreases to 41 gCO<sub>2</sub>e/km, which is 5.9 times less than its petrol equivalents.

In the worst-case scenario, the battery would be produced in China and the EV would run on the EU's most carbon intensive grid (Poland). In this situation, the lifetime BEV impact increases to 151 gCO<sub>2</sub>e/km, which would still be 37% cleaner than the petrol counterpart. This is an improvement compared to T&E's previous estimate (28%), partly explained by the fact that more recent data is used for China's electricity grid (see Annex 3).

In 2030, thanks to rapidly decarbonising electricity, the lifetime BEV impact decreases down to 46 gCO<sub>2</sub>e/km for an average EU medium electric car (4.6 times cleaner than that of a petrol car).



Source: T&E LCA analysis of a medium-sized car, battery assumed to be produced with the EU27 average grid and BEV charged with the average EU27 electricity grid.

Figure 2 – BEV lifecycle emissions savings in T&E updated model

## 2.2 Influence of national electricity grids over the use phase

The carbon intensity of the electricity used to charge the vehicle over its life has the highest impact over the lifecycle CO<sub>2</sub> emissions of electric cars. The results in this section correspond to a situation where the battery is produced with EU average electricity, so only the electricity used to charge the vehicle changes. The 2022 electricity grid carbon intensity of each country is defined based on a linear growth between 2020 data derived from EMBER and 2025 from ENTSO-E ‘National Trends’ scenarios. If an EV is recharged in Poland, which has the most carbon intensive electricity grids in Europe, an EV is still 40% cleaner than a petrol car (see Figure 3 below). In Germany, Italy, the Netherlands and Belgium, an electric car is slightly more than 2.5 times cleaner than a conventional car while EV in France and Sweden have impacts more than five times smaller (around 45 gCO<sub>2</sub>e/km).

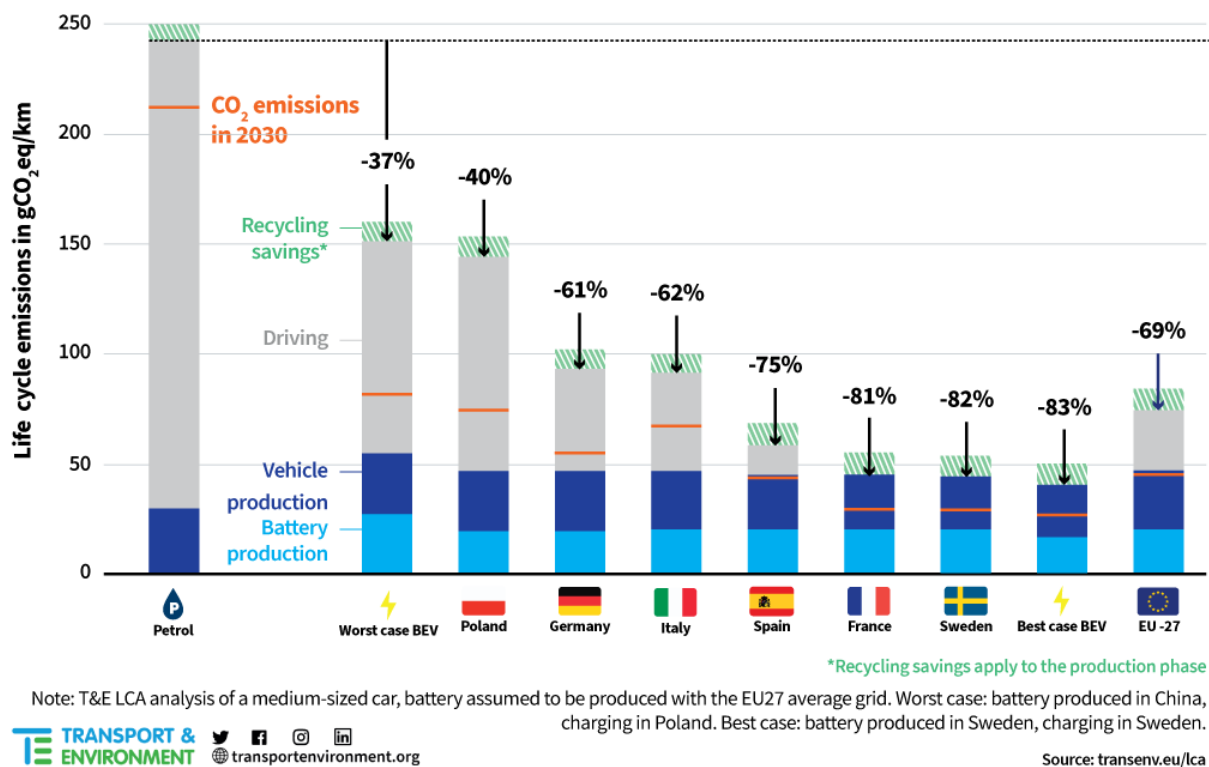


Figure 3 – Lifecycle emissions in key European countries

## 2.3 Lifecycle emissions of hybrid vehicles

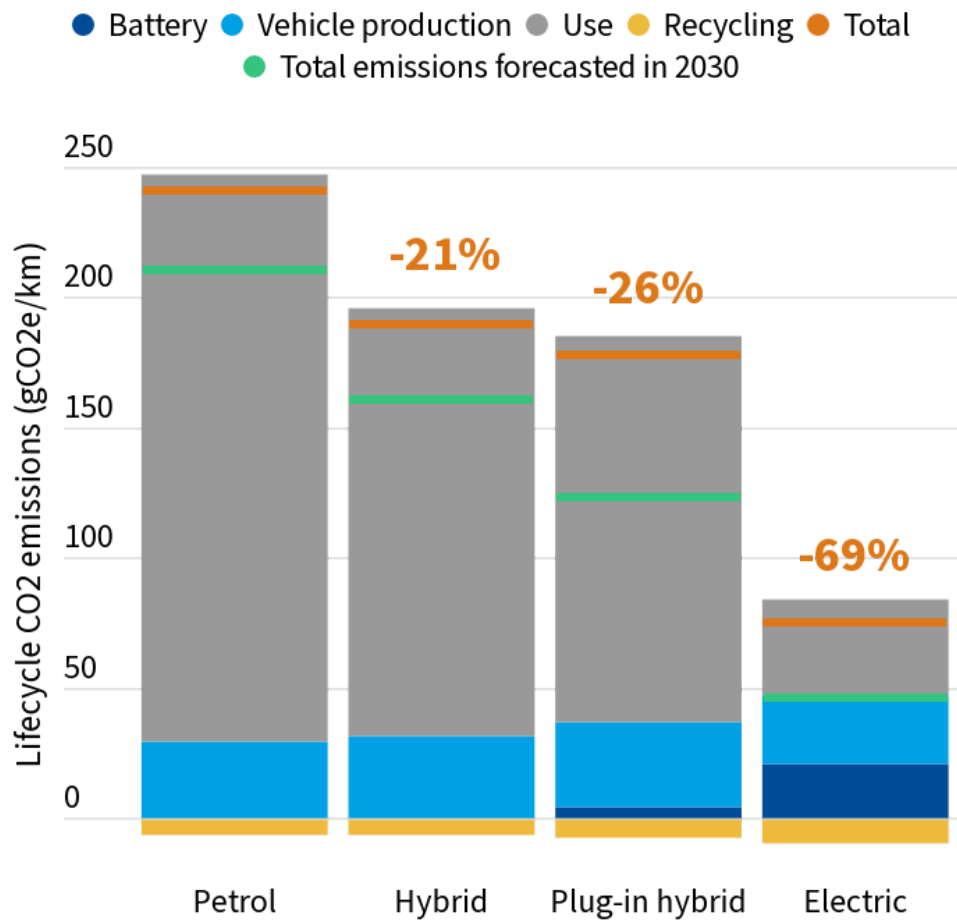
As described in section 1.2, HEVs and PHEVs were included in the T&E LCA model. The results show that the HEV powertrain used in a medium-sized car solely achieved a 21% reduction in LCA emissions compared to its petrol equivalent. For large and executive categories, the reductions are 23% and 16%<sup>22</sup> respectively. Regarding PHEVs<sup>23</sup>, their improvements compared to petrol cars are limited to 26% for medium cars (32% for large vehicles and 25% for executive vehicles). This shows that the lifecycle emissions of an average HEV and PHEV are much closer to conventional petrol engines and fall short of the emissions reduction achieved in BEVs. For instance, PHEVs have 2.4 times larger lifecycle emissions than BEVs for medium-sized cars.

For cars bought in 2030, all powertrain types benefit from LCA emissions reductions due to improvements in fuel and electricity consumption patterns. HEVs bought in 2030 are expected to be 23% cleaner than their petrol equivalents. The efficiency and utility factors of PHEVs are expected to improve such that their lifecycle emissions are projected to be 41% lower than petrol vehicles. But even with these

<sup>22</sup> The lower reduction in the case of the executive segment is explained by the relatively lower consumption from executive petrol cars where mild hybridisation already enables significant savings (mild hybrids are included in the “petrol” powertrain while HEV refers to full hybrids).

<sup>23</sup> As reported by ICCT in <https://theicct.org/publication/ghg-benefits-incentives-ev-mar22/>, there are large variations of LCA emissions between different PHEV models. T&E methodology outlined in this report is applicable to an average PHEV for each segment used by an average user. This methodology does not pretend to be accurate for specific models or specific user behaviour.

improvements, the average PHEV real world utility factor would still be limited to around 56%. As such, PHEV lifecycle emissions are still expected to be 2.7 times larger than those of the BEVs footprint, since the latter are able to fully benefit from grid decarbonisation.



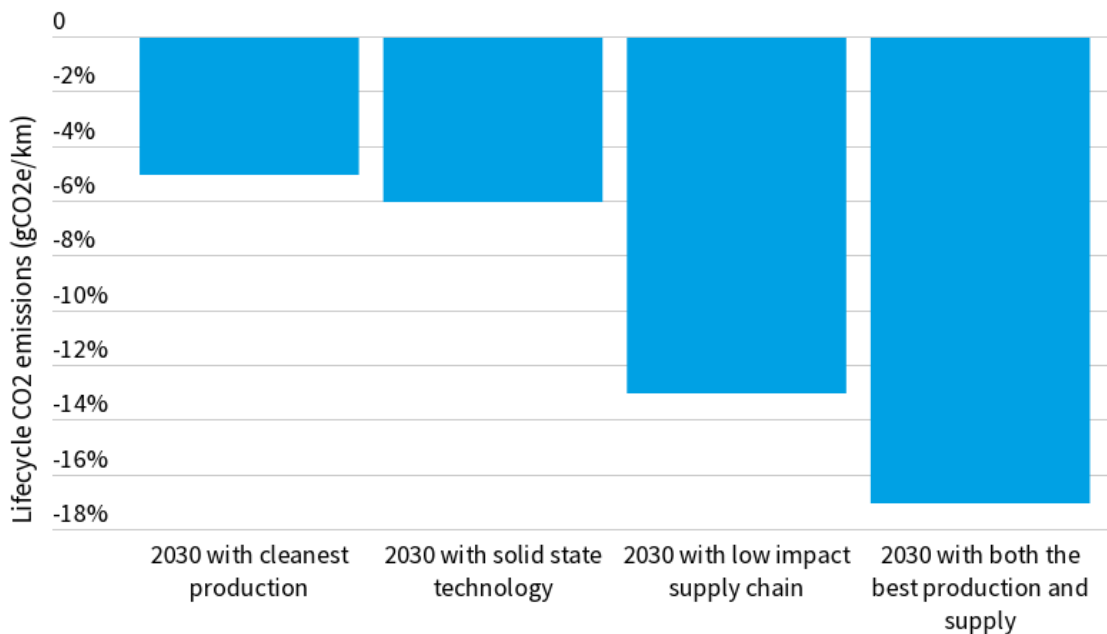
**Source:** T&E LCA analysis of a medium-sized car, battery assumed to be produced with the EU27 average grid, BEV/PHEV charging with the EU27 average grid

**Figure 4 – Lifecycle emissions per powertrain**



## 2.4 Sensitivity analyses on battery supply-chain, technology and production

This section analyses the impact of changes in different battery parameters now available in the LCA tool. The baseline is a medium-sized BEV charged with the EU average electricity grid from 2030 and an NMC-811 battery produced using the EU average grid; this baseline for cars bought in 2030 is already 39% cleaner than the average BEV bought in 2022. First, the results displayed in Figure 6 show that the use of the cleanest country electricity grid (Sweden) for production would enable a 5% decrease of the lifecycle emissions. Then, if a new technology such as solid-state batteries is used, the lifecycle emissions could decrease by 6% compared to the baseline. Finally, the parameter which has the highest influence on the results is the production route of battery materials: with best-in-class supply chains used for the sourcing of lithium, nickel or graphite, the lifecycle emissions of BEV could be decreased by 13%. With the lowest production impact where a LIB battery is produced in Sweden and a low impact supply chain, the lifecycle emissions of BEVs would be reduced by 17% compared to the baseline.



**Baseline:** Medium-sized BEV with battery produced in 2030 with EU27 average electricity grid, charged with the EU27 average grid, battery chemistry NMC-811. LCA emissions: 46 gCO<sub>2</sub>e/km

**Sensitivities:** - Cleanest battery production with the Swedish electricity grid  
 - SSB technology with a NMC811 cathode, lithium metal anode and oxide based solid electrolyte  
 - Low impact supply chain cumulating best practices for battery mineral extraction

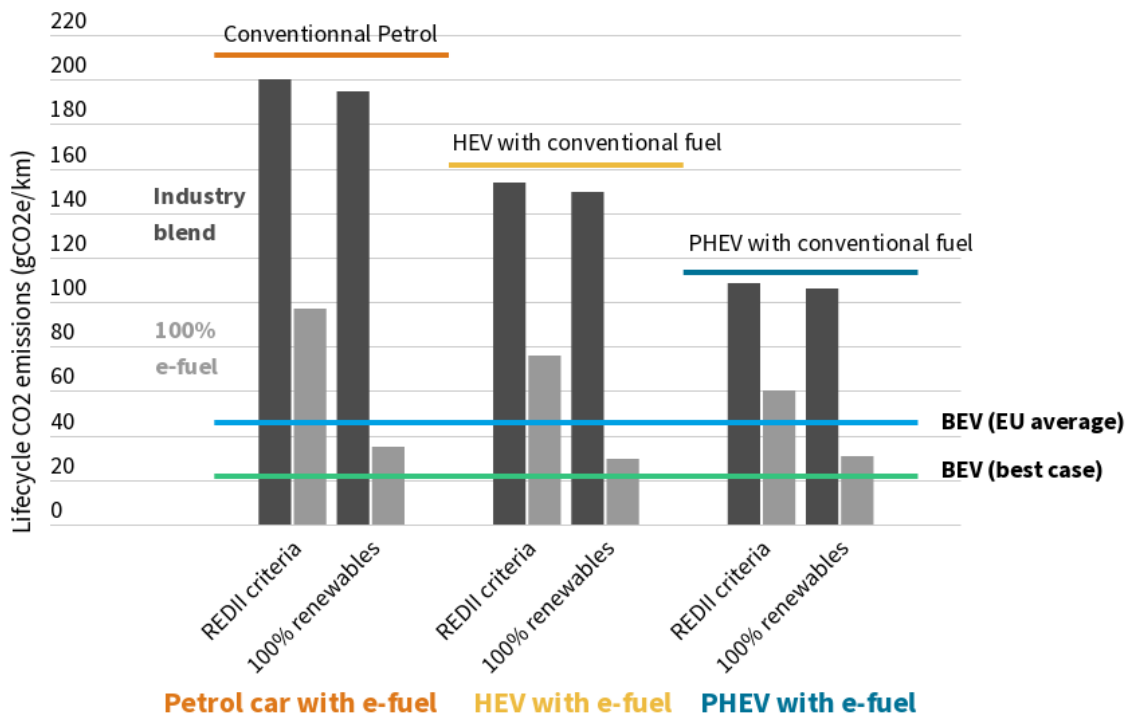
**Source:** T&E LCA model and battery carbon footprint based on a battery LCA by Minviro

**Figure 6 – BEV lifecycle emissions’ sensitivity to battery parameters**

In the most ambitious conditions where the lowest production and supply chain impacts are combined with the cleanest electricity grid in the use phase, the lifecycle emissions of BEV could become as low as 21 gCO<sub>2</sub>e/km.

## 2.5 Life cycle emissions of cars running on e-fuels

Lifecycle emissions of cars running on e-fuels depend highly on the energy source used to produce the fuel and the blending of fuel types in the final fuel available at the pump as shown in the different options in Figure 5. In the base case (“Industry blend”), the blending is chosen to be representative of the average blending of fuel in the market (according to the fuel industry forecast of the e-fuel production dedicated to road transport). When the e-fuel is produced according to the REDII criteria and blended with conventional petrol, the lifecycle emissions of a petrol car would only be reduced by 5% compared to conventional fuel (5% for HEV and 4% for PHEV). In the scenario when the e-fuel is produced from 100% renewables, this reduction reaches 8% for medium petrol cars and 7% for HEVs and PHEVs.



**REDII criteria:** the electricity used to produce the e-fuel is based on a 70% CO2 reduction of the WTW fuel emissions.  
**100% renewables:** based on the forecast of renewable energy sources used by electrolyzers.  
**BEV (best case):** Swedish electricity grid used for battery production and charging, and low supply-chain impacts.  
**Source:** T&E LCA analysis of medium-sized car bought in 2030

**Figure 5 – Lifecycle emissions of cars running on e-fuels for cars bought in 2030**

In the theoretical scenario where a petrol car runs on pure e-fuel produced with 100% renewable electricity, its lifecycle emissions would be 20% lower than a BEV charged with the EU average electricity mix. Nevertheless, if the e-fuel is produced in the best conditions from renewables, a fair comparison would need to consider the cleanest conditions for BEV production. In this best scenario for BEV where the cleanest production and supply chain are used (see section 2.4), BEV lifecycle emissions are 27%

lower than the best case for combustion engines where a HEV runs with pure e-fuel made from 100% renewable energy<sup>24</sup>. The main explanation for this difference is the low efficiency of the e-fuel production. Based on a 55% overall efficiency to convert electricity into e-fuel, a HEV running on e-fuel would need much more electricity compared to the direct use of electricity in a BEV, and so it generates a much greater amount of indirect emissions. This shows that BEVs have the highest potential to improve emissions over their lifecycle for cars bought in 2030.

Even in the best conditions, a car running solely on e-fuels is not cleaner through its lifecycle compared to a BEV running on the cleanest electricity grid. More realistic blending forecasts from the industry show that the expected low quantity of e-fuel would only enable marginal savings.

## Further information

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<sup>24</sup> Due to lower overall energy efficiency and higher production carbon footprint, PHEVs running on the best e-fuel and with the cleanest battery production have slightly higher lifecycle emissions than HEVs running on the best e-fuel.

# Annex

## 1. Emissions factors - hypotheses used

As the UNECE provides different values depending on technologies, the following assumptions were made:

- For coal, most power plants today use “pulverised coal (PC) technology” which is also the type with the highest carbon intensity. It is assumed that 80% of EU coal generation uses PC technology, with the remaining 20% being other technologies mentioned by UNECE (subcritical, supercritical and integrated gasification combined cycle).
- Carbon dioxide capture and storage (CCS) is not included for any electricity source.
- For solar photovoltaic (PV), the overwhelming majority of panels are polycrystalline silicon according to UNECE. It is assumed that 90% of photovoltaic panels are silicon-based (the average between rooftop-mounted and ground-mounted options from UNECE is used) and that 10% consist of other new technologies (thin-film PV).
- Regarding hydroelectricity, as stated by UNECE, the 360 MW plant is considered as the most representative.
- The offshore wind emissions factor is the average between the “concrete foundation” and “steel foundation” options provided by UNECE.

## 2. PHEV methodology

The lifecycle analysis of PHEVs relies on their real world fuel and electricity consumption data. The average real world fuel and electricity consumption can be calculated based on PHEVs consumption in different driving modes (charge depleting - CD, and charge sustaining - CS) and the share of the distance driven on electricity (UF).

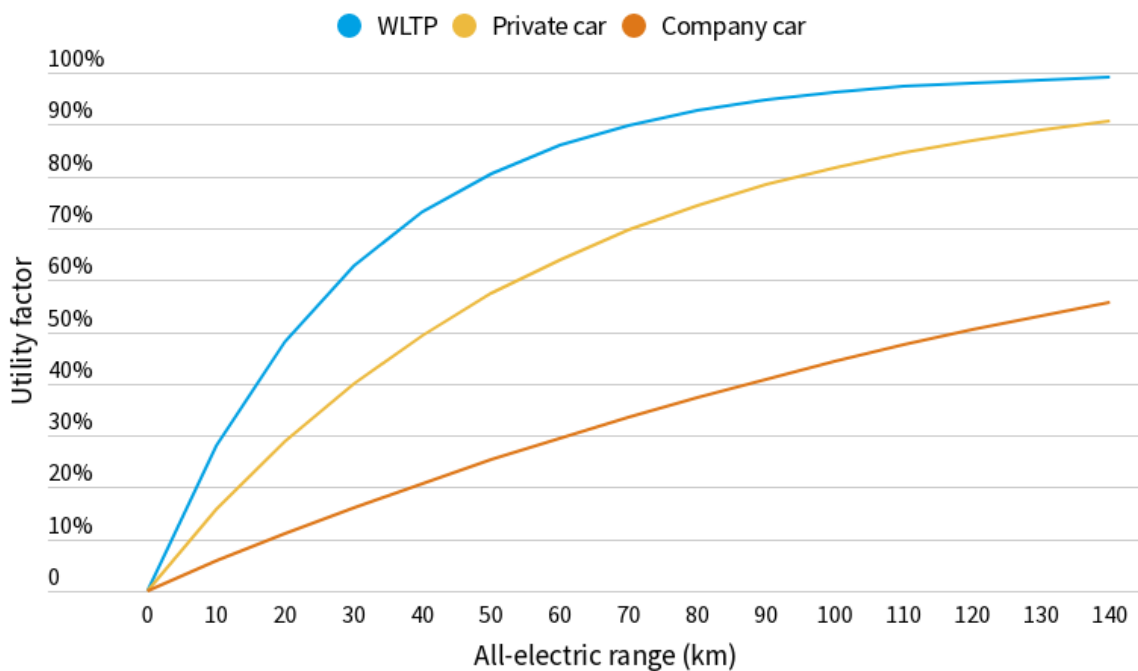
Data shows that the real world fuel consumption of PHEVs is two to four times higher than their type-approval values. The gap between real world and type approval values can mainly be explained by overly optimistic UF in the WLTP regulation compared to the real world performance of PHEVs. Moreover, the WLTP utility factor does not discriminate between private and company users. T&E has developed a methodology to derive the real world consumption of PHEVs.

As inputs, T&E methodology relies on data made available by carmakers on their websites for the latest PHEV models: carmakers usually provide fuel and electricity consumption and the Equivalent All Electric Range (EAER), as determined according to the WLTP. For cars where the electricity consumption is not advertised, the values are extracted from the EEA dataset<sup>25</sup>. T&E extracted data from carmakers' websites and the EEA for a large number of models. The sales volumes of the chosen models amounts to 81% of volume of 2021 EU PHEV sales. The methodology is as follows:

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<sup>25</sup> <http://co2cars.apps.eea.europa.eu/>

1. The WLTP theoretical utility factor is calculated based on the WLTP EAER.
2. The real world estimates of the fuel consumption in CS and CD mode are derived from the input data (WLTP average fuel and electricity consumption) and the WLTP UF.
3. Additional correction factors are included to convert WLTP fuel and electricity consumption into real-world estimates.
4. The utility factor of PHEVs in real-world driving conditions is derived from Fraunhofer ISI's UF real world estimates<sup>26</sup>, as shown in Figure 7.
5. Finally, knowing the estimate of all fuel and electricity consumption in the different phases (CD and CS) and the real world utility factor, the average real world fuel and electricity consumption can be derived.



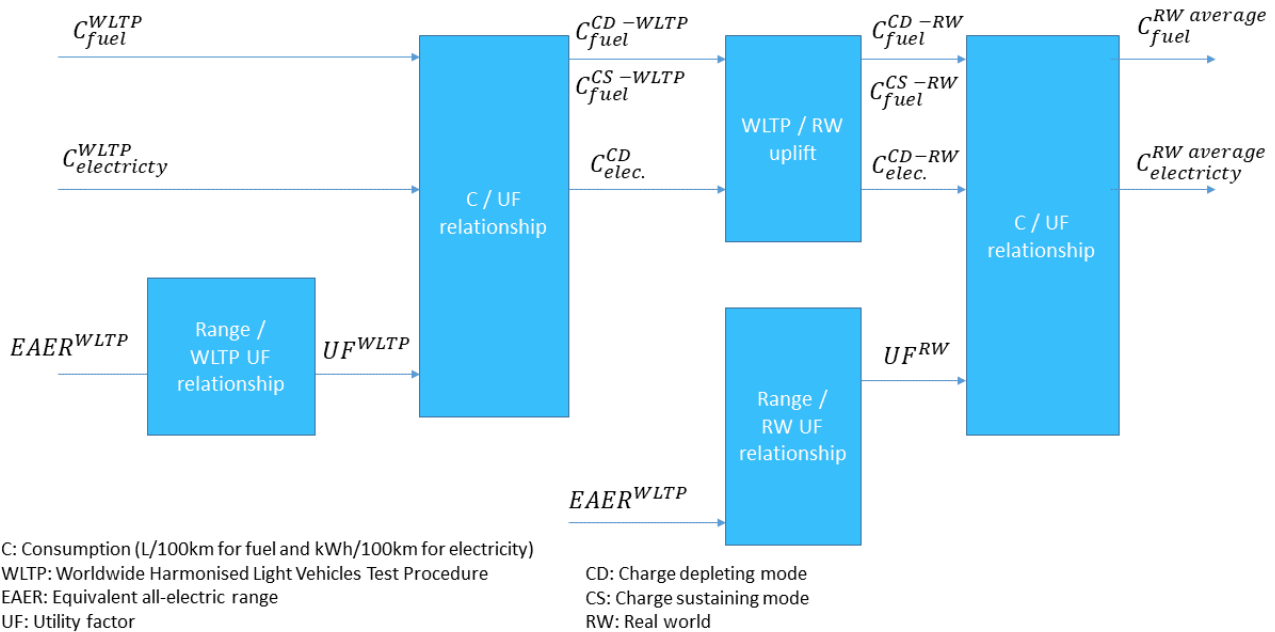
**Source:** Plötz, P. and Jöhrens, J. (2021): Realistic Test Cycle Utility Factors for Plug-in Hybrid Electric Vehicles in Europe. Karlsruhe: Fraunhofer Institute for Systems and Innovation Research ISI

**Figure 7 – Fraunhofer ISI’s realistic UF curve for private and company cars**

The outline of the method is depicted in Figure 8 below.

<sup>26</sup> [https://www.isi.fraunhofer.de/content/dam/isi/dokumente/cce/2021/BMU\\_Kurzpapier\\_UF\\_final.pdf](https://www.isi.fraunhofer.de/content/dam/isi/dokumente/cce/2021/BMU_Kurzpapier_UF_final.pdf)

Fraunhofer ISI used the UF function from the WLTP and fitted it to empirical PHEV usage for 1,385 private PHEV from Germany and 10,872 company cars from Germany and the Netherlands to define representative UF functions for both company and private cars.



**Figure 8 – Real-world PHEV fuel and electricity consumption calculation methodology**

The detailed methodology with each calculation steps is described below:

### 1. WLTP utility factor calculation

In the WLTP methodology, the utility factor is defined based on the range in charge depleting mode until the end of the transition cycle (R\_CDC). The transition cycle is the WLTC where the transition from charge depleting to charge sustaining modes is considered to have taken place. The R\_CDC is not publicly available; however, it may be estimated based on the WLTP equivalent all electric range (EAER) provided by carmakers. The EAER is defined as the part of the total range in charge-depleting mode that can be attributed to the use of the electric energy from the battery.

The R\_CDC and EAER differ because the test cycle carries on after the vehicle reaches its EAER so that the distance driven at the end of the transition cycle is larger than the EAER. In this report, T&E assumes that, on average, the EAER is reached during the transition cycle<sup>27</sup>. So, for the purpose of this analysis, the R\_CDC is assumed to be the distance reached at the end of the test cycle in which the EAER is reached. With this hypothesis, the WLTP utility factor can only take discrete values defined from the R\_CDC as shown in Table 5 (each test cycle of the WLTP has a fixed distance of 23 km). For instance, for a PHEV with an advertised EAER of 55km, we assume that the test cycle 3 is the transition cycle and we use a UF of 84%.

<sup>27</sup> Depending on PHEV models, the EAER can be reached during the transition cycle or in the preceding cycle. It is assumed that the average PHEV model reaches the EAER during the transition cycle.

**Table 5 – WLTP utility factor value depending on the test cycle**

WLTC	Range in charge depleting mode until the end of transition cycle (km)	UF
1	23	51%
2	47	73%
3	70	84%
4	93	89%
5	116	92%
6	140	94%

## 2. WLTP consumptions calculation

Having determined an approximation for the utility factor, the WLTP fuel and electricity consumption in CS and CD mode can be derived from the WLTP combined consumptions using the following formula:

$$C^{WLTP\ combined} = UF^{WLTP} \cdot C^{CD-WLTP} + (1 - UF^{WLTP}) \cdot C^{CS-WLTP}$$

Where:

$C^{WLTP\ combined}$  = weighted consumption (L/100km for fuel and kWh/100km for electricity) from carmakers data

$UF^{WLTP}$  = WLTP utility factor as a function of the range in CD mode  $R_{CDC}$

$C^{CD-WLTP}$  consumption in charge depleting mode

$C^{CS-WLTP}$  consumption in charge sustaining mode

In the WLTP, the value of the fuel consumption in charge-depleting mode is not generally zero as the ICE usually starts before the battery reaches a minimum state of charge and the end of the transition cycle. In order to simplify this formula, we assume that the fuel consumption in charge-depleting mode is about 10% of the consumption in charge sustaining mode. This assumption can be compared to results from the ADAC Ecotest reported by ICCT<sup>28</sup> for 9 vehicles. By using an average of the ratio of CD and CS consumption weighted by each vehicle's 2021 sales, T&E estimate that the fuel consumption in CD mode is about 15% of the consumption in CS mode in the ADAC Ecotest conditions. Nevertheless, the ADAC Ecotest expands the WLTC by an additional highway cycle and stronger accelerations in the transition. Therefore, the fuel consumptions (CD mode and CS mode) measured in these conditions may be overestimated compared to

<sup>28</sup> <https://theicct.org/publication/ghg-benefits-incentives-ev-mar22/>

WLTP test conditions. Therefore, a reduced 10% ratio was used as the ratio of CD mode fuel consumption and CS mode fuel consumption in the WLTP test to account for the lower average CD consumption in WLTP conditions compared to ADAC. This ratio is applied here to a theoretical average vehicle but individual vehicles can have a wide range of values. For instance, it ranges from only 7% for the Kia Optima Sportswagon to 32% for the BMW 225xe, based on values from ADAC reported by ICCT, the large difference can be attributed to the different carmaker designs of PHEVs.

The WLTP fuel consumption in CS mode can be estimated with the following formula:

$$C_{fuel}^{CS-WLTP} = \frac{C_{fuel}^{WLTP\ combined}}{1 - 90\% \cdot UF^{WLTP}}$$

The WLTP electricity consumption in CS mode is assumed to be equal to zero, as according to the regulation the state of charge of the battery may fluctuate during the charge sustaining test but overall all the vehicle should maintain the same state of charge. Therefore the value in CD mode can be calculated from the following formula:

$$C_{electricity}^{CD-WLTP} = \frac{C_{electricity}^{WLTP\ combined}}{UF^{WLTP}}$$

### 3. Real-world estimates from WLTP consumptions

The CD and CS mode fuel consumption figures in real-world driving conditions can be derived from their respective estimated WLTP value by applying a corrective factor. ICCT calculated that, on average, the real world fuel consumption of ICEs is 14% larger than the consumption achieved on the WLTP test<sup>29</sup>. T&E's methodology assumes that this relationship can also be applied to PHEVs. Regarding electricity consumption, T&E calculations from efficiency data of BEVs provided by the EV-Database<sup>30</sup> show that electricity consumption in real-world conditions is, on average, 5% higher than WLTP values. It is assumed that this relationship is also applicable for PHEVs used in CD mode.

<sup>29</sup>

<https://theicct.orgCO2e/publication/on-the-way-to-real-world-co2-values-the-european-passenger-car-market-in-its-first-year-after-introducing-the-wltp/>

<sup>30</sup> <https://ev-database.orgCO2e/>



#### 4. Real-world estimate of the utility factor

$$UF = 1 - \exp\left(-\sum_{i=1}^{10} c_i \left(\frac{AER}{d_n}\right)^i\right)$$

Where:

*AER* is the WLTP all-electric range in km

$c_i$  parameters are constants from the WLTP regulation (EC 2017)

The utility factor is defined based on the distance driven electrically and can therefore be described as a function of the all-electric range<sup>31</sup>. Fraunhofer defined a methodology to derive these UF curves based on real world data. This UF function is based on the official UF function from the WLTP regulation where the  $d_n$  parameter is scaled to fit empirical data using a standard statistical method. Fraunhofer ISI found that the best real world estimate leads to the following values of  $d_n$ :  $d_n = 1544$  km for private cars and  $d_n = 4500$  km for company cars<sup>32</sup>. As a comparison, the WLTP regulation uses  $d_n = 800$  km. Empirical data used by Fraunhofer ISI shows that private and company cars do not have the same utility factors. This difference is due to different charging and driving behaviour: company car drivers often own a fuel card from the employer and have no incentive to charge their car regularly and on average company cars drive more than double the km's of private cars.

Each PHEV car segment had a different share of private and company cars according to T&E analyses of 2021 registration data in EU27 provided by Dataforce. The share of private and company cars for each segment are provided in Table 6. Using this data the average UF can be calculated for each segment.

**Table 6 – Private / company car share in Europe**

	2021 company car share	2021 private share
C	62%	38%
D	81%	19%
E	83%	17%
All PHEVs	70%	30%

<sup>31</sup> Strictly speaking, the UF in the WLTP definition measures the share of km driven in charge depleting mode. Fraunhofer ISI expects their method of estimating the relationship between pure electric range and the share of km driven in pure electric mode to result in a very similar UF-curve compared to estimates based on the relationship between charge depleting range and share of km driven in charge depleting mode. However, Fraunhofer ISI outlined that more research on the relation between pure-electric UF and CD-mode UF is required.

<sup>32</sup> The 2022 update of these parameters carried-out by ICCT -not published at the time of writing- is expected to lead to a larger deviation compared to the WLTP regulation, meaning lower utility factors with more recent PHEV models compared to data reported by Fraunhofer ISI in April 2021.

An average UF can be calculated for each segment by weighting the UF of private and company cars by their respective share of sales in the given segment<sup>33</sup>. The larger share of company cars sold in segments D and E implies that the UF will be lower in those segments as the UF function of company cars has lower values than private cars (private car users tend to charge their cars more often than company car users and drive less). Average utility factor results are provided in Table 7. On average over all segments, the UF is estimated to be 37%.

## 5. Real-world average of fuel and electricity consumptions

The real-world estimate of the average fuel and electricity consumption can be derived thanks to the formula detailed in step 2, using the average real world UF of each segment and the real world consumption estimates calculated in step 3. In that case, we assume that the real-world fuel consumption in CD mode is about 15% of the fuel consumption in CS mode, as derived via the ADAC Ecotest results. Average fuel and electricity consumption values are provided in Table 8.

## 6. Projection for 2030

T&E’s model assumes that PHEV fuel consumption in CS and CD mode will decrease by 1.2% per year in line with our forecast for HEV, as described in section 1.2 and that the electricity consumption would decrease by 0.5% per year (in line with our hypothesis for BEV). In addition the UF of PHEVs in 2030 is modelled to increase for two reasons:

- Based on [T&E analysis](#) of carmakers production plans and the trend of the PHEV battery capacity increase, the electric range of PHEVs is expected to increase by 40% in 2030 compared to 2021.
- Following historical trends, the share of company cars among the new sales of PHEVs is expected to decrease in the future. For instance, 91% of PHEVs sales were company cars in 2015, but this share decreased to 77% in 2019, and again to 70% in 2021. Based on a linear forecast, T&E assumes that 50% of new PHEVs sales will be company cars in 2030.

This leads to the following change in the UF by segment:

**Table 7 – Real-world utility factor per segment**

	2022 real world UF	2030 real-world UF
C	40%	57%
D	33%	51%
E	37%	57%

<sup>33</sup> This analysis is applicable for the average new car of each segment and is not representative of specific individual user behaviour for a given model.

Segment C shows the largest utility factor, since it has the highest share of private cars. The company car share is higher in segments D and E, but segment E PHEVs have larger batteries and longer electric ranges<sup>34</sup>, so it has a higher UF than segment D. In 2030, the combination of the longer electric range and larger share of private cars results in an increase of the UF in each segment. On average over all segments, the UF would increase from 38% in 2022 to 56% in 2030.

The final fuel and electricity consumption used in the LCA model are shown in Table 8:

**Table 8 – Fuel and electricity consumption per segment**

	Average fuel consumption (L/100km)		Average electricity consumption (kWh/100km)	
	2022	2030	2022	2030
C	4.5	3.2	7.8	10.6
D	5.2	3.6	8.2	12.3
E	5.3	3.5	11.4	16.6

### 3. Location of battery production

The life cycle carbon intensity of the electricity grid options used for battery manufacturing are the following:

**Table 9 – LCA carbon intensity of electricity used for battery production in gCO<sub>2</sub>e/kWh**

	2022	2030
EU average	245	160
Sweden	35	29
Germany	355	250
China	680	580

The EU cleanest option is based on the electricity grid from Sweden. The option with higher-than-average carbon intensity is now based on the German grid. China's electricity grid life cycle carbon intensity is

<sup>34</sup> The WLTP range of cars in each segments are: 57km for segment C, 55km for segment D and 68km for segment E.

based on the IEA's Stated Policy (STEPS) scenario<sup>35</sup>. This result is a significant improvement as the 2020 value was assumed to be 919 gCO<sub>2</sub>e/kWh in T&E's LCA previous version based on Knobloch et al. (2020)<sup>36</sup>.

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<sup>35</sup> T&E estimates from values reported by ICCT in <https://theicct.orgCO2e/publication/a-global-comparison-of-the-life-cycle-greenhouse-gas-emissions-of-combustion-engine-and-electric-passenger-cars/>  
<sup>36</sup> <https://doi.org/10.1038/s41893-020-0488-7>